

November 15, 2018

Comments from Academics, Scientists and Clinicians on A Working Approach for Identifying Potential Candidate Chemicals for Prioritization

Submitted online via *Regulations.gov* to docket EPA-HQ-OPPT-2018-0659

These comments are submitted on behalf of the undersigned academics, scientists, and clinicians. We declare collectively that we have no direct or indirect financial or fiduciary interest in any chemical or product that is the subject of these comments. The co-signers' institutional affiliations are included for identification purposes only and do not imply institutional endorsement or support unless indicated otherwise.

We appreciate the opportunity to provide written comments on approaches for EPA's prioritization of existing chemicals pursuant to the Toxic Substances Control Act (TSCA), as amended by the Frank R. Lautenberg Chemical Safety for the 21st Century Act ("amended TSCA"). EPA held a meeting on prioritization on December 11, 2017 in Washington, D.C., which several of the undersigned attended. We followed up with detailed written comments in January 2018.¹ EPA is now requesting comment on its short and long-term plans for the prioritization process.²

There are over 40,000 chemicals on the active TSCA inventory.³ The most important outcomes of EPA's prioritization process are the identification, expeditious evaluation, and limiting of dangerous chemicals, leading to cleaner air, safer water, and safer products-- and ultimately, healthier lives for families. To accomplish this, EPA must ensure the prioritization process: (1) gathers sufficient information using the resources and data available; (2) applies the most current scientific principles and methods; and (3) is transparent, unbiased and accessible for all stakeholders.

Our comments address the following issues related to these points:

- 1. To gather sufficient data on chemicals for prioritization, EPA must do more than public notification for information gathering. It should exercise its full authorities to request testing and data from industry under TSCA sections 4, 8, and 14.**
- 2. EPA should not use "Application of systematic review in TSCA risk evaluations" in the prioritization process because the protocol and data quality criteria are arbitrary and not science-based.**
- 3. All stakeholders must have equal opportunities and support to participate in the prioritization process; additionally any opportunities for stakeholder participation must be made public and it is**

¹ UCSF PRHE, et al. (2018) Comments from Academics, Scientists, and Clinicians on Approaches for Identifying Potential Candidates for Prioritization for Risk Evaluation Under Amended TSCA. Available at: <https://prhe.ucsf.edu/sites/prhe.ucsf.edu/files/wysiwyg/2018%201%2025%20UCSF%20comments%20on%20pre-prioritization%20FINAL.pdf>.

² US EPA (2018) A Working Approach for Identifying Potential Candidate Chemicals for Prioritization. Available at: https://www.epa.gov/sites/production/files/2018-09/documents/preprioritization_white_paper_9272018.pdf

³ US EPA. (2018). *TSCA Chemical Substance Inventory* | US EPA. Available at: <https://www.epa.gov/tsca-inventory>.

inappropriate for any stakeholder to have special opportunities not available to other stakeholders.

- 4. Industry stakeholders wishing to sponsor development of information related to potential low-priority chemicals should pay a fee for EPA evaluation of the chemical's data to understand the gaps; EPA should then provide a detailed description of the information that the industry sponsor would need to develop to complete the database on the chemical.**

- 5. The long-term proposal of binning the entire TSCA inventory is unnecessary, resource-intensive and scientifically problematic. Instead, EPA should expand its TSCA workplan methodology, as this is both a sound approach for identifying high priority chemicals and will make the best use of Agency resources.**

We are appreciative of the opportunity to provide public input. Please do not hesitate to contact us with any questions regarding these comments.

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DETAILED COMMENTS

- 1. To gather sufficient data on chemicals for prioritization, EPA must do more than public notification for information gathering. It should exercise its full authorities to request testing and data from industry under TSCA sections 4, 8, and 14.⁴**

Given the dearth of toxicity data on most chemicals, EPA needs to exercise its full authority under TSCA to collect necessary data to evaluate chemicals for potential low and high priority designation. EPA can only designate a chemical as low priority if there is sufficient evidence to conclude that the chemical does not pose an unreasonable risk to health or environment, including to highly exposed or susceptible populations, as detailed below. Despite the critical need for data, EPA's only plan for generating information beyond what is already reasonably available is that it "may" issue public notifications regarding data gaps/ insufficient information on chemicals.⁵

Public notification will be insufficient to generate the specific data EPA will need. After its review of the data landscape for potential high and low priority chemicals, EPA should describe the key areas where data is lacking, and issue orders pursuant to TSCA Section 4 and/ or Section 8 to obtain these data. Section 4 test orders should outline the most relevant test models, exposure pathways, health outcomes, and target populations (including any vulnerable or sensitive populations) anticipated to support the generation of high-quality and relevant evidence to support timely decision-making, as described in more detail below.

The responses and data received from EPA requests should be publicly available. TSCA section 14 clearly states that health and safety studies are not considered confidential business information (CBI) and thus are not protected from disclosure. EPA should also provide a public summary characterizing the data and its completeness for each chemical.

- 2. EPA should not use "Application of systematic review in TSCA risk evaluations" in the prioritization process because the protocol and data quality criteria are arbitrary and not science-based.**

When screening information on potential high-priority chemicals, EPA states that "The initial emphasis will be the exclusion of unacceptable data sources based on data quality criteria outlined in the Application for Systematic Review in TSCA Risk Evaluations EPA document."⁶

This is inappropriate because the data quality criteria and 'scoring' method in the TSCA systematic review document are incompatible with the best scientific methods and approaches. We detail why the TSCA systematic review method is invalid, scientifically flawed, and should not be used in the comments we submitted on the systematic review method to EPA on July 26, 2018 (see attached Appendix).

- 3. All stakeholders must have equal opportunities and support to participate in the prioritization process; additionally any opportunities for stakeholder participation must be made public and it is inappropriate for any stakeholder to have special opportunities not available to other stakeholders.**

⁴ U.S.C. §2603, §2607 and §2613, referred to in these comments as TSCA 4, 8 and 14.

⁵ US EPA (2018) A Working Approach for Identifying Potential Candidate Chemicals for Prioritization. Pp. 9, 15.

⁶ Id. Pp. 13.

EPA states that “stakeholders” suggested they could sponsor the development of information on additional low priority chemicals, and that there may be “an enhanced stakeholder role” in designation of these substances.⁷ The stakeholders referred to are clearly industry, as industry are the only stakeholders with the resources to sponsor the development of information on chemicals.

First, it is inappropriate for ANY stakeholder to play an ‘enhanced’ role in EPA’s decision-making on high or low-priority determinations. The Agency alone should make this determination. Stakeholders can, of course, provide information to EPA to inform its decision.

Second, to ensure transparency and an unbiased decision, EPA must publicly announce opportunities for stakeholder participation and EPA should provide sufficient access and resources for participation of all relevant stakeholders, including those among the public, to participate.

4. Industry stakeholders wishing to sponsor development of information related to potential low-priority chemicals should pay a fee for EPA evaluation of the chemical’s data to understand the gaps; EPA should then provide a detailed description of the information that the industry sponsor would need to develop to complete the database on the chemical.

To designate substances as low priority, the Agency must find that they do not present unreasonable risks to health or the environment, and it must do so based on “sufficient information” by law.

The concept of “sufficient information” covers both the types of data needed and the quality of each piece of data. For industry stakeholders who would like to sponsor development of information, EPA should clearly define what constitutes “sufficient information” for evaluation to make a low priority designation. This definition should include a list of traits deemed important to assess, such physical characteristics, health outcomes, effects on potentially exposed or susceptible subpopulations, etc. and a discussion of how these traits will be evaluated to determine whether “sufficient information” is available.

Some health hazard datasets EPA has utilized which could inform the Agency’s definition include:

- The dataset from studies that determine hazard to humans required by EPA’s Office of Pesticide Programs for pesticide registration⁸
- The health hazard dataset needed for EPA’s Design for the Environment (DfE) program to conduct an alternatives assessment⁹
- The eleven health and ecological hazard endpoints listed in Figure 6 of the prioritization working document¹⁰

The interested industry stakeholder(s) should pay a fee for EPA to develop “completeness metrics” that track how many of the desired traits could be assessed based on the available data and provide a public summary characterizing the “completeness of the database” for the chemical(s) of interest. EPA has adopted similar approaches in the past, for instance using published criteria from the HPV, Chemical

⁷ Id. Pp. 17.

⁸ <https://www.epa.gov/pesticide-registration/data-requirements-pesticide-registration>. Accessed Jan 15, 2018.

⁹ US EPA (2011) Design for the Environment Alternatives Assessment Criteria for Hazard Evaluation. Available: https://www.epa.gov/sites/production/files/2014-01/documents/aa_criteria_v2.pdf

¹⁰ US EPA (2018) A Working Approach for Identifying Potential Candidate Chemicals for Prioritization. Pp. 25.

Assessment and Management Program (ChAMP) and EPA’s Risk Assessment Guidance to evaluate the data adequacy in its brominated phthalates Data Needs Assessment.¹¹

Where data are lacking, EPA should explicitly outline for the industry sponsor(s) the studies needed, including descriptions of the relevant test models, exposure pathways, health outcomes, and target populations (including any vulnerable or sensitive populations).

To describe such data gaps in a consistent and comprehensive manner, EPA needs to clearly define the data needed to determine that a chemical does not pose particular hazards—i.e., is not a carcinogen, is not a developmental or reproductive toxicant, etc. EPA’s definition should be consistent with established approaches of other agencies such as the National Toxicology Program (NTP),¹² the International Agency for Research on Cancer (IARC),¹³ and EPA’s own guidelines including the Cancer Guidelines.¹⁴

These guidelines explicitly define what constitutes the determination of no hazard. For example, in the EPA Cancer Guidelines, a determination of “Not Likely to Be Carcinogenic to Humans” requires **robust evidence** as follows:

“This descriptor is appropriate when the available data are considered robust for deciding that there is no basis for human hazard concern. In some instances, there can be positive results in experimental animals when there is strong, consistent evidence that each mode of action in experimental animals does not operate in humans. In other cases, there can be convincing evidence in both humans and animals that the agent is not carcinogenic. The judgment may be based on data such as:

- animal evidence that demonstrates lack of carcinogenic effect in both sexes in well-designed and well-conducted studies in at least two appropriate animal species (in the absence of other animal or human data suggesting a potential for cancer effects),
- convincing and extensive experimental evidence showing that the only carcinogenic effects observed in animals are not relevant to humans,
- convincing evidence that carcinogenic effects are not likely by a particular exposure route (see Section 2.3), or
- convincing evidence that carcinogenic effects are not likely below a defined dose range.

A descriptor of “not likely” applies only to the circumstances supported by the data. For example, an agent may be “Not Likely to Be Carcinogenic” by one route but not necessarily by another. In those cases that have positive animal experiment(s) but the results are judged to be not relevant to humans, the narrative discusses why the results are not relevant.”¹⁵

¹¹ US EPA (2015) TSCA Work Plan Chemical Problem Formulation and Data Needs Assessment: Brominated Phthalates Cluster Flame Retardants. Pg. 7. Available at: https://www.epa.gov/sites/production/files/2015-09/documents/brominated_phthalates_cluster_data_needs_assessment.pdf

¹² National Toxicology Program, Office of Health Assessment and Translation. 2015. Handbook for Conducting a Literature-Based Health Assessment Using OHAT Approach for Systematic Review and Evidence Integration. Available from: https://ntp.niehs.nih.gov/ntp/ohat/pubs/handbookjan2015_508.pdf

¹³ IARC (International Agency for Research on Cancer). 2006. IARC Monographs on the Evaluation of Carcinogenic Risks to Humans: Preamble. Available: <http://monographs.iarc.fr/ENG/Preamble/index.php>

¹⁴ US EPA (2005) Guidelines for Carcinogen Risk Assessment. Available from: https://www.epa.gov/sites/production/files/2013-09/documents/cancer_guidelines_final_3-25-05.pdf

¹⁵ US EPA (2005) Guidelines for Carcinogen Risk Assessment. Pg. 84-85. Available from: https://www.epa.gov/sites/production/files/2013-09/documents/cancer_guidelines_final_3-25-05.pdf

For example, if an industry sponsor has only negative genotoxic data and SAR (structure-activity relationship) data related to a chemical's carcinogenicity, these are not sufficient to conclude that the chemical is not carcinogenic according to the EPA guidelines. Based on the guidelines, supporting data from male and female animals of at least two species in well-designed and conducted studies would be required. This is the type of directions EPA would need to provide to industry sponsors to guide their development of information.

In general, "New Approach Methods" (NAM) data alone cannot be used to designate a chemical as low-priority as this is not consistent with the statute's requirement for "sufficient information." A recent National Academies of Sciences (NAS) report describes how data generated by NAMs (such as high-throughput screening methods) are not different than any other type of *in vitro* or cell-based assay data that would be considered in reviewing evidence on a chemical.¹⁶ These kinds of assays provide mechanistic data, and the NAS report explicitly considered how mechanistic data could be utilized in evidence integration. The committee came to two conclusions. First, a pre-defined protocol for evaluating relevance and study quality must be developed for the mechanistic data evidence stream, as is done for other evidence streams such as epidemiological or toxicology data. This must include the pre-defined determinations of what types of evidence will be included/excluded, how the risk of bias and quality of studies will be evaluated, and how the mechanistic evidence will be integrated with other evidence streams. These protocols should be made publicly available for stakeholder review prior to the initiation of the assessment. As an example, in the NAS' case study on phthalates, the committee was not able to integrate results from high-throughput assays because the cell lines used were of unknown relevance to the *in vivo* mechanism of phthalate toxicity.¹⁷

Finally, the foundation of a hazard classification is the animal and human data, with the mechanistic data playing a supporting role. If mechanistic data is relevant, it can be used to upgrade a hazard classification, or increase the confidence of a finding made based on evaluation of animal and human data. A hazard classification is never made based on high-throughput or other kinds of NAM data alone.^{18,19} EPA should continue to follow this established scientific practice when evaluating data submitted by industry sponsors.

5. The long-term proposal of binning the entire TSCA inventory is unnecessary, resource-intensive and scientifically problematic. Instead, EPA should expand its TSCA workplan methodology, as this is both a sound approach for identifying high priority chemicals and will make the best use of Agency resources.

EPA proposes to "bin" the entire TSCA inventory through calculating a score based on hazard, exposure and information availability data. The binning process proposed is complex and resource-intensive, requiring information gathering on close to 40,000 chemicals. First, this process is unnecessary. As we

¹⁶ The National Academies of Sciences (2017) Application of Systematic Review Methods in an Overall Strategy for Evaluating Low-Dose Toxicity from Endocrine Active Chemicals. Washington, D.C.: National Academies Press.

¹⁷ The National Academies of Sciences (2017) Application of Systematic Review Methods in an Overall Strategy for Evaluating Low-Dose Toxicity from Endocrine Active Chemicals. Washington, D.C.: National Academies Press. Pg. 78

¹⁸ Id. Pp. 10; 158-9

¹⁹ NTP (2015) Handbook for Conducting a Literature-Based Health Assessment Using OHAT Approach for Systematic Review and Evidence Integration

recommended in our January 2018 comments, EPA should instead augment its already-established TSCA workplan methodology.²⁰ Second, we are concerned about the scientific approaches in the binning methodology.

Expansion of the TSCA Work Plan Methodology

The TSCA Work Plan Methodology was developed through a transparent process with multiple opportunities for stakeholder participation, and it is informed by current scientific principles, reliable data sources and priority public health considerations.

EPA should build on the strengths of the Workplan methodology and augment it to meet the needs of amended TSCA.

To ensure enough chemicals in the prioritization pipeline, EPA should expand the universe of chemicals feeding into Step 1 of the workplan methodology. We recommend:

- Including polymers and mixtures. GreenScreen recently developed hazard criteria for polymers and mixtures which EPA could draw upon.²¹
- Expanding hazard endpoints considered. Chronic toxicity, respiratory toxicity/ sensitization, immunotoxicity, and endocrine activity could all be included.
- Expanding authoritative list sources to include: the CA Safer Consumer Products candidate chemical list;²² CA Proposition 65 carcinogens, reproductive & developmental toxicant; all chemicals requiring Toxics Release Inventory reporting; Hazardous Air Pollutants, and ATSDR neurotoxicants.
- Including additional exposure criteria such as increases in production volume, storage near water, and/ or detection in surface, ground, or drinking water. These track well with factors required to be considered in prioritization by amended TSCA.²³
- Including chemicals that are of concern to sub-populations with greater exposure and/ or susceptibility: EPA should engage in dialogues with workers, tribes, environmental justice and fence-line communities (those located in proximity to the conditions of use of TSCA chemicals) to determine the chemicals of highest concern to these populations, and ensure they are included in Step 1.

Amended TSCA mandates broader consideration of vulnerable populations, and EPA should expand the criteria considered in the Workplan method. To address populations with greater exposures and susceptibilities, EPA should integrate considerations on environmental justice populations, such as by incorporating data from EPA's EJ Screen tool into a chemical's score.²⁴

²⁰ UCSF PRHE, et al. (2018) Comments from Academics, Scientists, and Clinicians on Approaches for Identifying Potential Candidates for Prioritization for Risk Evaluation Under Amended TSCA.. Available at: <https://prhe.ucsf.edu/sites/prhe.ucsf.edu/files/wysiwyg/2018%201%2025%20UCSF%20comments%20on%20pre-prioritization%20FINAL.pdf>.

²¹ Clean Production Action (2018) GreenScreen for Safer Chemicals Hazard Assessment Guidance For Chemicals, Polymers and Products. Available: https://www.greenscreenchemicals.org/images/ee_images/uploads/resources/GreenScreen_Guidance_v1_4_2018_01_Final.pdf

²² <https://www.dtsc.ca.gov/SCP/CandidateChemicals.cfm>

²³ Requirements in U.S.C. §2605(b)

²⁴ US EPA (2018) EJSCREEN: Environmental Justice Screening and Mapping Tool. Available at: <https://www.epa.gov/ejscreen>

Finally, a goal of any final prioritization process should be to put chemicals with insufficient information into the pipeline for data development. EPA only says that missing data will be “flagged for potential future information gathering.” The prioritization strategy needs to have an explicit process for selecting which chemicals with insufficient information will be followed up, and then utilizing EPA’s Section 4 and 8 information gathering authorities to generate the needed data.

Problems with the binning approach

EPA’s binning approach combines toxicity and exposure information into ratios (HER, BER or TER). This is not appropriate because the ratio relies heavily on exposure potential considerations based on current uses and production. Production volumes and uses of chemicals can and do rapidly change. Separately scoring hazard and exposure domains is a more robust approach, as EPA did in the TSCA Workplan methodology and demonstrated in Candidate Scoring Methods 1, 2, 3 and 4 in its Discussion Document from 2017.²⁵ Further, as described below, these ratios (HER, BER, TER) likely do not reflect chemical’s true risk.

Of the 4 methods that separately score hazard and exposure, some incorporate NAM data. A comparison of the outputs of these methods for scoring chemicals as high, moderate or low concern with a method using the hazard/ bioactivity exposure ratio (H/BER) is shown in the Figure below. The H/BER method stands out from the other 4 by ranking significantly fewer chemicals as potentially high priority and significantly more chemicals as potentially low priority. This suggests that the hazard to exposure ratio analysis is flawed and does not reflect the true risk of chemicals. EPA should not use methods that combine hazard and exposure information; instead it should separately score hazard and exposure domains.

²⁵ Discussion Document: Possible Approaches and Tools for Identifying Potential Candidate Chemicals for Prioritization. (2017). Pp. 54-57. Available at: https://www.epa.gov/sites/production/files/2017-11/documents/final_pre-prioritization_-_discussion_document_11.13.17.pdf

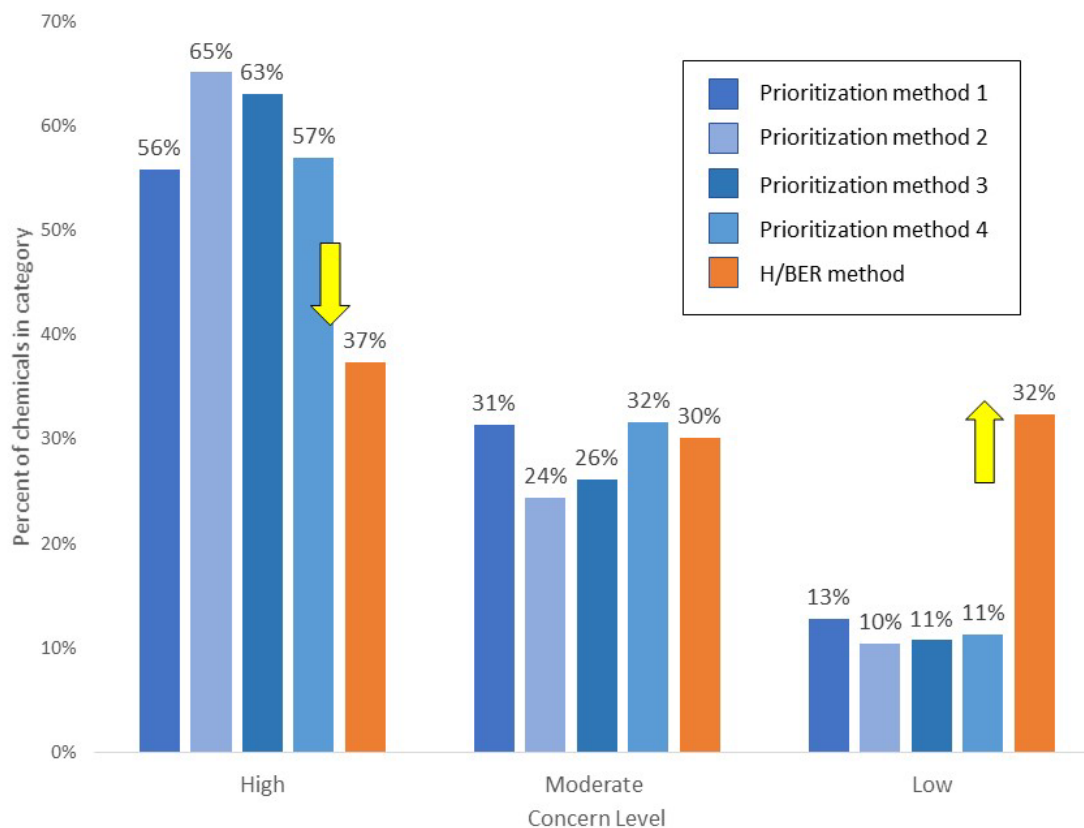


Figure: The method incorporating the hazard to bioactivity exposure ratio (H/BER) ranks significantly fewer chemicals as potentially high concern, and significantly more chemicals as potentially low concern. Data from EPA Discussion Document²⁶.

²⁶ Discussion Document: Possible Approaches and Tools for Identifying Potential Candidate Chemicals for Prioritization. (2017). Pp. 61, 63. Available at: https://www.epa.gov/sites/production/files/2017-11/documents/final_pre-prioritization_-_discussion_document_11.13.17.pdf

Appendix

Comments from Academics, Scientists and Clinicians on: The Application of Systematic Review in TSCA
Risk Evaluations

August 16, 2018

August 16, 2018

Comments from Academics, Scientists and Clinicians on: The Application of Systematic Review in TSCA Risk Evaluations.

Submitted online via *Regulations.gov* to docket EPA-HQ-OPPT-2018-0210

These comments are submitted on behalf of the undersigned academic, scientists, and clinicians. We declare collectively that we have no direct or indirect financial or fiduciary interest in any chemical under consideration in these risk evaluations. The co-signers' institutional affiliations are included for identification purposes only and do not necessarily imply any institutional endorsement or support, unless indicated otherwise.

We appreciate the opportunity to provide written comments on the Application of Systematic Review in TSCA Risk Evaluations,^a pursuant to the Toxic Substances Control Act (TSCA), as amended by the Frank R. Lautenberg Chemical Safety of the 21st Century Act (Lautenberg TSCA). TSCA requires that EPA make decisions about chemical risks based on the "best available science" and the "weight of the scientific evidence"^b which EPA defined in regulation as "...a systematic review method, applied in a manner suited to the nature of the evidence or decision, that uses a pre-established protocol to comprehensively, objectively, transparently, and consistently identify and evaluate each stream of evidence, including strengths, limitations, and relevance of each study and to integrate evidence as necessary and appropriate based upon strengths, limitations, and relevance."^c

Systematic review methods originated more than 40 years ago in psychology. The methodology was soon adapted to evaluating the effectiveness of clinical interventions in medicine and related disciplines in response to empirical evidence demonstrating the need to apply scientific principles not only to primary research, but also to research synthesis methods that inform decision-making in healthcare (1-3). Almost a decade ago, these empirically-proven methods for research synthesis were adapted to environmental health (4, 5). To date, science-based methods for systematic review in environmental health have been demonstrated in case studies in the peer-reviewed literature (6-13), and adopted by the National Toxicology Program (14) and the U.S. EPA's Integrated Risk Information System (IRIS) program (15).

EPA's systematic review framework under TSCA establishes EPA's "rules" for assembling and interpreting the scientific evidence on chemicals in commerce. These "rules" will determine, whether explicitly, implicitly, and/or by default, *what* evidence EPA will consider, and *how* it will evaluate that evidence when it is making decisions about potentially hazardous chemicals in commerce. Exposure to industrial, commercial, and consumer product chemicals is ubiquitous from the time of conception until death. As such, EPA's rules for gathering and interpreting the science that evaluates the relationship between these exposures and adverse health effects are of profound importance to the general public, and will have even greater impact on the potentially exposed or susceptible sub-populations Congress explicitly mandated EPA to protect: pregnant women, children, individuals with underlying health conditions, workers, and those with greater exposure and/or greater vulnerability to chemical toxicity and exposure.

^a 83 FR 26998, June 11, 2018

^b 15 USC §2625 (h)-(i)

^c 40 CFR 704.33

With so much at stake, we are deeply concerned by EPA's ad hoc and incomplete TSCA systematic review framework, which is inconsistent with current, established, best available empirical methods for systematic review. Moreover, as we detail below, the application of EPA's TSCA framework would likely result in the exclusion of quality research from EPA's decision-making. Accordingly, the TSCA systematic review method does not meet the mandate of the law to use the "best available science."^d

Based on the most current empirically demonstrated principles of systematic review methods, we provide EPA with concrete recommendations and approaches to correct its methodology and inform timely science-based decision-making to achieve the Agency's mission of protecting the public from harmful chemicals.

Our comments address the following six main points:

1. EPA's TSCA systematic review framework is ad hoc, incomplete, and does not follow established methods for systematic review that are based on the best available science.

We recommend: EPA should implement a systematic review method that is compatible with empirically based existing methods and aligns with the Institute of Medicine's^e definition of a systematic review, including but not limited to, using explicit and pre-specified scientific methods for every step of the review. EPA should consider methods demonstrated for use in environmental health, and which have been endorsed and utilized by the National Academy of Sciences, i.e., the National Toxicology's Office of Health Assessment and Translation systematic review method, and the Navigation Guide Systematic Review Method. EPA's TSCA systematic review framework should be peer-reviewed by qualified external experts in the field.

2. EPA's TSCA systematic review framework utilizes a quantitative scoring method that is incompatible with the best available science in fundamental ways:

- a. **Quantitative scores for assessing the quality of an individual study are arbitrary and not science-based; the Cochrane Collaboration and National Academy of Sciences recommend against such scoring methods.**
- b. **EPA's scoring method wrongly conflates how well a study is reported with how well the underlying research was conducted; and**
- c. **EPA's scoring method excludes research based on one single reporting or methodological limitation.**

We recommend: EPA should not use a quantitative scoring method to assess quality in individual studies; it should not conflate study reporting with study quality; and it should not exclude otherwise quality research based on a single reporting or methodological limitation. Rather EPA should employ a scientifically valid method to assess risk of bias of individual studies.

3. EPA's TSCA systematic review framework does not consider financial conflicts of interest as a potential source of bias in research.

^d 15 USC §2625 (h)

^e The Institute of Medicine is now the National Academy of Medicine.

We recommend: EPA should assess study and author funding source as a risk of bias domain for individual studies.

- 4. The literature review step of EPA's TSCA systematic review framework incorporates select best practices, but also falls short of, or is unclear about, many other best practices for conducting a systematic and transparent literature review.**

We recommend: EPA should make its framework for conducting a literature review congruent with all of the Institute of Medicine's best practices and explicitly include rules for when the list of relevant studies will be considered final.

- 5. EPA's TSCA systematic review framework correctly recognizes that mechanistic data are not required for a hazard assessment, but EPA is not clear that these data, if available, can only be used to increase, and not to decrease, confidence in a body of evidence.**

We recommend: EPA should be explicit that mechanistic data can only be used to upgrade a hazard classification, or increase the confidence of a finding made based on evaluation of animal and human data, and that these data will not be used to decrease confidence in a body of evidence.

- 6. EPA's TSCA systematic review framework is not independent of the regulatory end user of the review.**

We recommend: EPA's TSCA systematic reviews should be produced independently of the regulatory end user of the review.

We are appreciative of the opportunity to provide public input. Please do not hesitate to contact us with any questions regarding these comments.

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DETAILED COMMENTS

1. EPA's TSCA systematic review framework is ad hoc, incomplete, and does not follow established methods for systematic review that are based on the best available science.

The best available scientific method for a systematic review (SR) specifies that all components of a review be established in a publically available protocol written *prior* to conducting the review to minimize bias and to ensure transparency in decision-making. For example, the Institute of Medicine defines a systematic review as a “scientific investigation that focuses on a specific question and uses explicit, prespecified scientific methods to identify, select, assess, and summarize the findings of similar but separate studies” (emphasis added) (16)(p.1). A fatal flaw in EPA's SR framework is that it lacks essential SR elements, including but not limited to: (1) a protocol for executing a SR developed *prior* to conducting the SR; (2) an explicit method for evaluating the overall body of each evidence stream, i.e., animal, human, etc.; and (3) an explicit method for integrating two or more streams of evidence, including defined criteria for the type and level of evidence needed for a decision by EPA.

Notably, EPA's TSCA SR Framework presents a diagram of a complete SR framework in Figure 3-1 (page 15) and states in footnote 4 on that page that the:

Diagram depicts systematic review process to guide the first ten TSCA risk evaluations. It is anticipated that the same basic process will be used to guide future risk evaluations with some potential refinements reflecting efficiencies and other adjustments adopted as EPA/OPPT gains experience in implementing systematic review methods and/or approaches to support risk evaluations within statutory deadlines (e.g., aspects of protocol development would be better defined prior to starting scoping/problem formulation).

However, EPA's TSCA SR Framework then proceeds to describe an ad hoc and highly flawed method limited to only the data collection and, to a limited extent, the data evaluation components of a SR. Specifically, Figure S-1 below, excerpted from the National Academy of Sciences 2014 review of the EPA IRIS program's systematic review method (17), presents all of the components of a science-based SR. The red box indicates the parts of a SR method that EPA has included in its proposed framework.

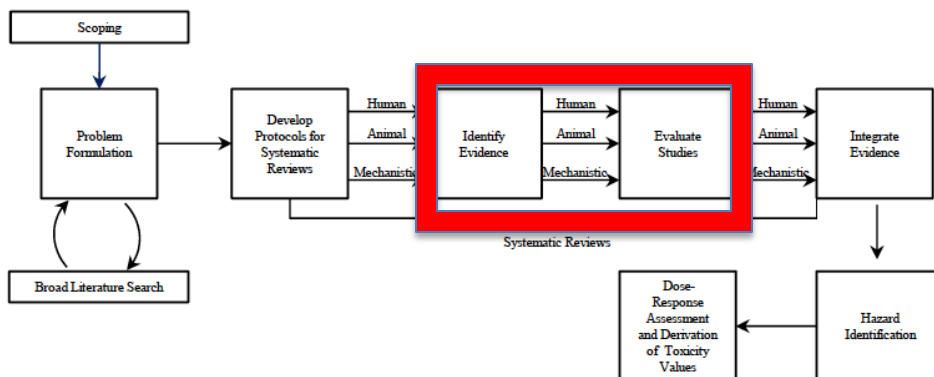


FIGURE S-1 Systematic review in the context of the IRIS process. The committee views public input and peer review as integral parts of the IRIS process, although they are not specifically noted in the figure.

EPA’s piecemeal approach is not only in direct contradiction with the best available scientific methods for SR, but also incompatible with the regulatory definition of “weight of evidence” in the risk evaluation rule, which specifies a complete method spelled out in a protocol developed *before* conducting the review. Therefore, the TSCA systematic review method violates both TSCA statute and regulation.^g

EPA explicitly states that it is proceeding with its first ten risk assessments in the absence of a pre-defined protocol and a complete method for systematic review. Specifically, EPA’s SR Framework states:

(p. 9) ... the purpose of the document is internal guidance that ... sets out general principles to guide EPA’s application of systematic review in the risk evaluation process for the first ten chemicals ... **EPA had limited ability to develop a protocol document detailing the systematic review approaches and/or methods prior to the initiation of the risk evaluation process for the first ten chemical substances. For these reasons, the protocol development is staged in phases while conducting the assessment work” (emphasis added).** Additional details on the approach for the evidence synthesis and integration will be included with the publication of the draft TSCA risk evaluations.

In effect, EPA is saying it does not have time to comply with its regulatory requirement to conduct a science-based systematic review, and will not actually develop its protocol until it completes the first ten systematic reviews.

First, this approach is in clear violation with scientifically-validated approaches to conducting systematic reviews. In its review of the EPA’s Integrated Risk Information System (IRIS) program’s proposed SR methods, the National Academy of Sciences specified that, “Completing the literature search as part of

^f EPA’s risk evaluation rule (40 CFR 704.33) states: “Weight of the scientific evidence means a systematic review method, applied in a manner suited to the nature of the evidence or decision, that uses a pre-established protocol to comprehensively, objectively, transparently, and consistently identify and evaluate each stream of evidence, including strengths, limitations, and relevance of each study and to integrate evidence as necessary and appropriate based upon strengths, limitations, and relevance.”

^g 15 USC §2625 (h)-(i) and 40 CFR 704.33

protocol development is inconsistent with current best practices for systematic review, and the IRIS program is encouraged to complete the public-comment process and finalize the protocol before initiating the systematic review” (15)(Pg. 8). In the case of TSCA risk assessments, EPA is not only completing the literature search as part of protocol development, it is completing the entire systematic review in the absence of a protocol and complete method. It is blatantly biased to write the rules of evidence assembly and interpretation at the same time one is applying the rules, and as such, this method cannot be validly referred to as a science-based systematic review.

Second, a lack of time is not a credible rationale for EPA’s failure to conduct a science-based systematic review for the first ten TSCA chemicals. There are multiple well-developed, science-based, peer-reviewed and validated methods for conducting systematic reviews in environmental health that EPA could readily apply, including the SR method and handbook developed by the Office of Health Assessment and Translation at the National Toxicology Program (14), and the Navigation Guide Systematic Review Method, which has been demonstrated in six case studies (6-13). The National Academy of Sciences cited both of these SR methods as exemplary of the type of methods EPA should use in hazard and risk assessment (17, 18). Further, the National Academy of Sciences utilized both methods in its 2017 assessment of the potential health impacts of endocrine active environmental chemicals (19). Specifically, in its 2017 review the National Academy of Sciences found:

The two approaches [OHAT and Navigation Guide] are very similar ... and they are based on the same established methodology for the conduct of systematic review and evidence assessment (e.g., Cochrane Collaboration, AHRQ Evidence-based Practice Center Program, and GRADE). Both the OHAT and Navigation Guide methods include the key steps recommended by a previous National Academies committee (NRC 2014) for problem formulation, protocol development, specifying a study question, developing PECO statement, identifying and selecting the evidence, evaluating the evidence, and integrating the evidence” (19)(page 119).

Protocols developed for applying the Navigation Guide and the OHAT method have been published and can serve as a template to further expedite EPA’s TSCA reviews.^h

Furthermore, the language of EPA’s systematic review framework is confusing, contradictory, and poorly and incorrectly referenced with little science or policy foundation. This suggests the authors of EPA’s TSCA Systematic Review Framework lack sufficient understanding of the scientific process integral to this work. A particularly egregious example is EPA’s stated understanding of EPA’s TSCA statutory science standards:

(Pg. 26) EPA/OPPT is required by TSCA to use the weight of the scientific evidence in TSCA risk evaluations. Application of weight of evidence analysis is an integrative and interpretive process that considers both data/information in favor (e.g., positive study) or against (e.g., negative study) a given hypothesis within the context of the assessment question(s) being evaluated in the risk evaluation.

This directly contradicts EPA’s own published rule which defines what a systematic review is (see

^h All Navigation Guide systematic review protocols can be found at: <https://prhe.ucsf.edu/navigation-guide> The National Toxicology Program’s protocol for its systematic review to evaluate the evidence for an association between exposure to PFOA or PFOS and immunotoxicity or immune-related health effects is at: https://ntp.niehs.nih.gov/ntp/ohat/pfoa_pfos/protocol_201506_508.pdf

footnote “e”, above) and such an understanding completely subverts the purpose of a systematic review which is to explicitly avoid a simplistic analysis that would lead to erroneous conclusions along the lines of stating that, for instance, “five studies are in favor (positive) and ten are against (negative) and therefore the weight is ...”

Another bewildering statement by EPA concerns its highly quantitative scoring method, which is the main topic of its systematic review framework (see comment #2, below). EPA adds a caveat to the scoring method that says quantitative scoring is actually a qualitative method, and further: “The [scoring] system is not intended to imply precision and/or accuracy of the scoring results” (Pg. 35).

The ad hoc and incomplete nature of EPA’s systematic review framework is incompatible in many additional fundamental ways, described further in detail below, with science based methods of systematic review developed, endorsed, and/or advanced by the: National Academy of Sciences (17-19); the Institute of Medicine (16); the National Toxicology Program (14); the Cochrane Collaboration (20); the Grading of Recommendations Assessment, Development and Evaluation (GRADE) method (21, 22); the international scientific collaboration that developed a framework for the “systematic review and integrated assessment” (SYRINA) of endocrine disrupting chemicals (23); the SYRCLE systematic review method for animal studies (24); the Campbell Collaboration’s methods (25); and the Navigation Guide systematic review method developed by a collaboration of scientists led by the University of California San Francisco (4). Most of these organizations also pre-publish their protocols either online (i.e., the National Toxicology Program) or in PROSPEROⁱ (i.e., UCSF).

We recommend: EPA should implement a systematic review method that is compatible with empirically based existing methods and aligns with the Institute of Medicine’s definition of a systematic review, including, but not limited to, using explicit and pre-specified scientific methods for every step of the review. EPA should consider methods demonstrated for use in environmental health, and which have been endorsed and utilized by the National Academy of Sciences, i.e., the National Toxicology’s Office of Health Assessment and Translation systematic review method, and the Navigation Guide Systematic Review Method. EPA’s TSCA systematic review framework should be peer-reviewed by qualified external experts in the field.

ⁱ PROSPERO International prospective register of systematic reviews <https://www.crd.york.ac.uk/prospero/>

2. EPA's TSCA systematic review framework utilizes a quantitative scoring method that is incompatible with the best available science in fundamental ways:

- a. Quantitative scores for assessing the quality of an individual study are arbitrary and not science-based; the Cochrane Collaboration and National Academy of Sciences recommend against such scoring methods.**
- b. EPA's scoring method wrongly conflates how well a study is reported with how well the underlying research was conducted; and**
- c. EPA's scoring method excludes research based on one single reporting or methodological limitation.**

A detailed explanation of each of these scientific shortcomings is provided below.

(a) Quantitative scores for assessing the quality of an individual study are arbitrary and not science-based.

EPA's SR framework employs a quantitative scoring method to assess the quality of individual studies, assigning, based on its "professional judgment", various weights for quality domains and then summing up the quantitative scores to decide whether a study is of "high", "medium", or "low" quality as follows:^j

(Pg. 33) A numerical scoring method is used to convert the confidence level for each metric into the overall quality level for the data/information source. The overall study score is equated to an overall quality level (*High*, *Medium*, or *Low*) using the level definitions and scoring scale shown in Table A-1. The scoring scale was obtained by calculating the difference between the highest possible score of 3 and the lowest possible score of 1 (i.e., $3-1=2$) and dividing into three equal parts ($2 \div 3 = 0.67$). This results in a range of approximately 0.7 for each overall data quality level, which was used to estimate the transition points (cut-off values) in the scale between *High* and *Medium* scores, and *Medium* and *Low* scores. These transition points between the ranges of 1 and 3 were calculated as follows: Cut-off values between *High* and *Medium*: $1 + 0.67 = 1.67$, rounded up to 1.7 (scores lower than 1.7 will be assigned an overall quality level of *High*) Cut-off values between *Medium* and *Low*: $1.67 + 0.67 = 2.34$, rounded up to 2.3 (scores between 1.7 and lower than 2.3 will be assigned an overall quality level of *Medium*)

This overall scoring method is applied to all streams of evidence, and our comments reflect our objection to EPA's applying scoring to any and all streams of evidence.^k

Illustrative of the scoring method, in Appendix H "Data Quality Criteria for Epidemiologic Studies," (page

^j See Appendix A for a more detailed description of the scoring method; how the method will be applied specifically to various streams of evidence, i.e., occupational exposure and release data; animal and in vitro data; epidemiologic studies; etc., is described in subsequent Appendices B-H.

^k EPA's framework applies quantitative scoring to all types of data; EPA/OPPT "is not applying weighting factors to the general population, consumer, and environmental exposure data types. In practice, it is equivalent to assigning a weighting factor of 1, which statistically assumes that each metric carries an equal amount of weight." (Pg. 96).

225) EPA presents how scoring is further applied to human studies, explaining:

The critical metrics within each domain are those that cover the most important aspects of the domain and are those that more directly evaluate the role of confounding and bias. After pilot testing the evaluation tool, EPA recognized that more attention (or weight) should be given to studies that measure exposure and disease accurately and allow for the consideration of potential confounding factors. Therefore, metrics deemed as critical metrics are those that identify the major biases associated with the domain, evaluate the measurement of exposure and disease, and/or address any potential confounding. ... EPA/OPPT assigned a weighting factor that is twice the value of the other metrics within the same domain to each critical metric. Remaining metrics are assigned a weighting factor of 0.5 times the weighting factor assigned to the critical metric(s) in the domain. The sum of the weighting factors for each domain equals one.

There is no scientific evidence to support EPA's selection of these "critical metrics" as being more important than other metrics, i.e., why within the "study participation" domain "selection" and "attrition" are more important than "comparison group"; and there are no data supporting EPA's choice of particular numbers for weighting these 'critical metrics' (i.e., some metrics are "twice" as important as the other metrics).

Overall, there is no scientific justification for EPA to assign these or any other quantitative scoring measures for assessing the quality of an individual study. The implicit assumption in quantitative scoring methods is that we know empirically how much each risk of bias domain contributes to study quality, and that these domains are independent of each other. This is not a scientifically supportable underlying assumption. Research has documented that scoring methods have, at best, unknown validity, may contain invalid items, and that results of a quality score are not scientifically meaningful or predictive of the quality of studies (26-28). An examination of the application of quality scores in meta-analysis found that quality-score weighting produced biased effect estimates, with the authors explaining that quality is not a singular dimension that is additive, but that it is possibly non-additive and non-linear (29).

Aggregating across quality criteria to produce a single score is recognized by preeminent systematic review methodologists as problematic and unreliable because the weights assigned are arbitrary and focus on the quality of reporting rather than the design and conduct of the research (21, 30). Scoring is not utilized by empirically based systematic review methodologies, such as the Cochrane Collaboration or GRADE (21, 31). As stated by the Institute of Medicine, "... systematic review teams have moved away from scoring systems to assess the quality of individual studies toward a focus on the components of quality and risk of bias" (16).

The Cochrane Collaboration, founded in 1993, is an international non-profit and independent organization that produces and disseminates systematic reviews of healthcare interventions and is a key locus of the world's most authoritative expertise on systematic review methods. Cochrane's methodology states: "The current standard in evaluation of clinical research calls for reporting each component of the assessment tool separately **and not calculating an overall numeric score (emphasis added)**"(31).

The National Academy of Sciences in its review of the EPA's IRIS program's method for SR, strongly supported a methodology that did not incorporate quantitative scoring, stating:

... Cochrane discourages using a numerical scale because calculating a score involves choosing a weighting for the subcomponents, and such scaling generally is nearly impossible to justify (Juni et al. 1999). Furthermore, a study might be well designed to eliminate bias, but because the study failed to report details in the publication under review, it will receive a low score. Most scoring systems mix criteria that assess risk of bias and reporting. However, there is no empirical basis for weighting the different criteria in the scores. Reliability and validity of the scores often are not measured. Furthermore, quality scores have been shown to be invalid for assessing risk of bias in clinical research (Juni et al. 1999). The current standard in evaluation of clinical research calls for reporting each component of the assessment tool separately and not calculating an overall numeric score (Higgins and Green 2008) (17)(Pg. 69).

b) EPA's scoring method wrongly conflates how well a study is reported with how well the underlying research was conducted.

Study reporting addresses how well research findings are written up, i.e., whether there is a complete and transparent description of what was planned, what was done, what was found, and what the results mean. Guidelines and checklists for authors have been developed to help ensure all information pertinent to assessing the quality and meaning of research is included in the report. The "Strengthening of Reporting of Observational Studies in Epidemiology" or "STROBE" Initiative is an example of a checklist of items that should be included in articles reporting such research.¹

EPA's SR Framework uses reporting measures in its scoring of the quality of human studies, including incorporating reporting guidelines into the reasons for scoring studies "low quality" (Metrics 1 and 15) or "unacceptable for use" (Metrics 2, 3, 4, 6, 7). EPA's SR Framework acknowledges that reporting is not the same as an underlying flaw in study methodology (Pg. 31), but then proceeds to ignore this distinction by using reporting as a measure of the quality of the underlying research. EPA's SR Framework not only does not "untangle" reporting from quality, it specifically conflates the two by using metrics in the STROBE reporting guidelines to score individual studies. The authors of the STROBE guidelines specifically note the guidelines are not a measure of the quality of the underlying research, stating:

The STROBE Statement is a checklist of items that should be addressed in articles reporting on the 3 main study designs of analytical epidemiology: cohort, case control, and cross-sectional studies. The intention is solely to provide guidance on how to report observational research well; these recommendations are not prescriptions for designing or conducting studies. Also, while clarity of reporting is a prerequisite to evaluation, the checklist is not an instrument to evaluate the quality of observational research (emphasis added). ... Our intention is to explain how to report research well, not how research should be done. We offer a detailed explanation for each checklist item. Each explanation is preceded by an example of what we consider transparent reporting. This does not mean that the study from which the example was taken was uniformly well reported or well done; nor does it mean that its findings were reliable, in the sense that they were later confirmed by others: it only means that this particular item was well reported in that study."(32)

How completely and clearly a study is reported is not a scientifically valid measure of the quality of the

¹ See Strobe statement at: <https://www.strobe-statement.org/index.php?id=strobe-aims>

underlying research (20, 21, 33, 34). As GRADE methodologists have succinctly stated, "... just because a safeguard against bias is not reported does not mean it was neglected"(21). Moreover, including many reporting items that are irrelevant to bias in a quality scoring rule (e.g., an indicator of whether power calculations were reported), will disproportionately reduce some of the resulting scores (29).

The Cochrane Collaboration Handbook for conducting a SR clearly distinguishes reporting and bias, the latter which is defined as "a systematic error, or deviation from the truth, in results or inferences" (20). The Cochrane Manual for conducting systematic reviews is explicit about not conflating reporting with bias, stating:

Bias may be distinguished from **quality**. The phrase 'assessment of methodological quality' has been used extensively in the context of systematic review methods to refer to the critical appraisal of included studies. The term suggests an investigation of the extent to which study authors conducted their research to the highest possible standards. This *Handbook* draws a distinction between assessment of methodological quality and assessment of risk of bias, and recommends a focus on the latter. The reasons for this distinction include:

1. The key consideration in a Cochrane review is the extent to which results of included studies should be *believed*. Assessing risk of bias targets this question squarely.
2. A study may be performed to the highest possible standards yet still have an important risk of bias. For example, in many situations it is impractical or impossible to blind participants or study personnel to intervention group. It is inappropriately judgemental to describe all such studies as of 'low quality', but that does not mean they are free of bias resulting from knowledge of intervention status.
3. Some markers of quality in medical research, such as obtaining ethical approval, performing a sample size calculation and reporting a study in line with the CONSORT Statement (Moher 2001d), are unlikely to have direct implications for risk of bias.
4. An emphasis on risk of bias overcomes ambiguity between the quality of reporting and the quality of the underlying research (although does not overcome the problem of having to rely on reports to assess the underlying research).

Importantly, in the application of EPA's SR Framework, studies can be scored as "low quality," and even excluded from EPA's review, based solely on a deficiency in reporting, irrespective of the quality of the underlying research. Research documents that important information is often missing or unclear in published research (35), as word limits, styles, and other specifications are highly variable, and non-standardized among peer-reviewed journals. As such, efforts to improve reporting are focused on uptake of reporting guidelines by journal editors and researchers (32, 36, 37). Improving reporting is needed in academic research, but as stated by the developers of the STROBE guidelines, "We want to provide guidance on how to report observational research well. ... the checklist is not an instrument to evaluate the quality of observational research."

Given the historical and present-day deficiencies in how studies are reported in the peer-reviewed literature, and because EPA's scoring system rates as 'unacceptable for use' any human study that does not report even one of five reporting metrics, EPA's proposal could reasonably be expected to lead to the exclusion from EPA's consideration much of the existing body of knowledge on the impact of

environmental chemicals on human health, and is inconsistent with TSCA mandates to use the “best available science” and “reasonably available information.”^m Applying flawed exclusion criteria that directly contradicts widely accepted empirically based SR methodological approaches will almost certainly result in flawed conclusions and threaten the protection of the public’s health.

(c) EPA’s scoring method excludes research based on one single reporting or methodological limitation.

In the “fatal flaw” component of EPA’s SR Framework’s scoring system, for each type of evidence stream, i.e., epidemiologic, animal, *in vitro*, etc., EPA created an arbitrary list of metrics that make studies “unacceptable for use in the hazard assessment,” stating:

EPA/OPPT plans to use data with an overall quality level of *High, Medium, or Low* confidence to quantitatively or qualitatively support the risk evaluations, but does not plan to use data rated as *Unacceptable*. Studies with any single metric scored as 4 will be automatically assigned an overall quality score of *Unacceptable* and further evaluation of the remaining metrics is not necessary (emphasis added). An *Unacceptable* score means that serious flaws are noted in the domain metric that consequently make the data unusable (or invalid) (Pg. 227).

There is no empirical basis for EPA’s selected list of fatal flaws.

Illustrative of this “fatal flaw” aspect of EPA’s scoring system, for human epidemiologic studies (See Section H.5, Table H-8 (page 231), EPA lists six domains of study quality, i.e., study participation; exposure characterization; outcome assessment; potential confounding/variable control; analysis; and other considerations for biomarker selection and measurement, and 19 metrics to assess the six domains. A study that has even one of the 19 “serious flaws” metrics is considered to be “unacceptable for use.”

The underlying assumptions of EPA’s “serious flaws” metrics are not science-based because:

- **EPA's list of "serious flaws" are not all equal indicators of study quality:**
For example, among human observational studies, any one of the list of 19 metrics can eliminate a study from consideration as EPA considers all of these "flaws" to be of equal import; as described in detail above, such weighting is arbitrary and not a science-based method.
- **EPA's list of "serious flaws" are not all related to real flaws in the underlying research:**
 - **Reporting** guidelines are wrongly equated with “serious flaws” in study quality. For example, in scoring the quality of human studies, 5 of 19 “serious flaw” metrics (Table H-8) are STROBE reporting guidelines (STROBE checklist items # 6,7,8,13,15). A study would be scored as “unacceptable for use” by EPA based on any one of these STROBE reporting guidelines. As described above in comment #2a, the STROBE guideline developers explicitly state this is neither the intended nor a scientifically valid use of these guidelines. (32)

^m 15 USC §2625(h) and (k)

- **Analysisⁿ** is equated with a "serious flaw" in study quality, but statistical power^o alone is not a valid measure of study quality. For example, EPA's framework excludes human studies that do not meet EPA's criteria for "high" in the analysis domain. EPA does not state how it will calculate whether a study is "adequately" powered. According to EPA's framework, to be included in an EPA review, a study must meet the "high" criteria in EPA's "Metric 13, Statistical power (sensitivity, reporting bias)" as presented in the box below. Studies that are not "high" quality for this metric would be designated as "unacceptable for use" by EPA:

Metric 13. Statistical power (sensitivity, reporting bias)
Instructions: To meet criteria for confidence ratings for metrics where 'AND' is included, studies must address both of the conditions where "AND" is stipulated. To meet criteria for confidence ratings for metrics where 'OR' is included studies must address at least one of the conditions stipulated.

EPA Metric 13. Excerpted from Table H-9 (page 243)

High (score = 1)	<p><u>For cohort and cross-sectional studies:</u> The number of participants are adequate to detect an effect in the exposed population and/or subgroups of the total population.</p> <p>OR</p> <p>The paper reported statistical power high enough ($\geq 80\%$) to detect an effect in the exposure population and/or subgroups of the total population.</p> <p><u>For case-control studies:</u> The number of cases and controls are adequate to detect an effect in the exposed population and/or subgroups of the total population.</p> <p>OR</p> <p>The paper reported statistical power was high ($\geq 80\%$) to detect an effect in the exposure population and/or subgroups of the total population.</p>
Medium (score = 2)	<ul style="list-style-type: none"> • Do not select for this metric.
Low (score = 3)	<ul style="list-style-type: none"> • Do not select for this metric.
Unacceptable (score = 4)	<ul style="list-style-type: none"> • <u>For cohort and cross-sectional studies:</u> The number of participants are inadequate to detect an effect in the exposed population and/or subgroups of the total population. • <u>For case-control studies:</u> The number of cases and controls are inadequate to detect an effect in the exposed population and/or subgroups of the total population.

ⁿ See Table H-8 "Serious Flaws that Would Make Epidemiological Studies Unacceptable for Use in the Hazard Assessment" under the "analysis domain" "statistical power/sensitivity" metric (page 233) "in conjunction with Table H-9 "Evaluation Criteria for Epidemiologic Studies, Metric 13 "statistical power (sensitivity, reporting bias) (page 243).

^o A power calculation is an estimate of the size of the study population needed to detect an effect of a given size.

First and foremost, EPA provides no method for how it will determine the “adequacy” of the statistical power of a study on which to base its score, and provides no rationale for excluding studies with less than 80% statistical power. According to STROBE guideline developers, ... “before a study is conducted power calculations are made with many assumptions that once a study is underway may be upended; further, power calculations are most often not reported” (32).

EPA’s Metric 13 statistical power/sensitivity also appears to confuse bias with imprecision. Individual studies that are “underpowered” (for example, because in the real world the exposed population may not be large enough for statistical purposes even if they are health impacted) can still be potentially valuable to science-based decision-making. For example a small study may be imprecise but that should not be confused with whether it is biased (20); a small study can be imprecise but at the same time less biased than a larger study (17). Small “underpowered” studies can also be combined in a meta-analysis that increases the statistical power of the body of evidence to reflect the relationship between an exposure and a health impact. Additionally, “underpowered” studies that find a health effect to be present may be indicative of a larger effect size than anticipated. Thus, omitting such studies would severely bias the conclusions of the review.

Illustrative of how EPA’s “analysis” metric could result in excluding high quality research that can inform science-based decision-making by EPA, in a 2017 systematic review by Lam et al. “*Developmental PBDE Exposure and IQ/ADHD in Childhood: A Systematic Review and Meta-analysis*,” (12) none of the 4 high-quality^p studies included in the meta-analysis reported a power calculation, and yet together, these studies found “a 10-fold increase (in other words, times 10) in PBDE exposure associated with a decrement of 3.70 IQ points (95% confidence interval:0.83,6.56).” It is also notable that one of the studies in the meta-analysis, Herbstman et al. 2010, (38) was assessed by the review authors to be “probably high risk of bias” for “Incomplete Outcome Data.”^q As such, this otherwise high quality study, i.e., all of the other domains were “definitely” or “probably” low risk of bias, would meet EPA’s criteria for “unacceptable for use” based on STROBE reporting guideline #15, “Report numbers of outcome events or summary measures over time”.^r

In short, the *Lam et al* systematic review, using the best available scientific methods, found that a ubiquitous environmental contaminant is impacting human intelligence, a finding that was subsequently reviewed and endorsed by the National Academy of Sciences (19). Yet EPA’s SR review framework would exclude crucial pieces of this body of evidence based on the Agency’s inaccurate, non-science-based criteria for deeming studies ‘unacceptable.’ This is contrary to TSCA’s mandate to use the best available science.^s

- **"Level of exposure" is equated with a "serious flaw".**

^p “High quality” defined as “definitely” or “probably” low or very low risk of bias (Figure 2a in the *Lam et al* paper) based on specific and detailed definitions of risk of bias established before the review was conducted.

^q The authors of the systematic review rated the Herbstman 2010 study “probably high risk of bias” for “incomplete outcome data” based on the following rationale: “Concerns regarding missing outcome data at each follow-up time on almost half the cohort of 210 with cord blood PBDE measurements; no argument is presented that would invalidate the possibility of a selection bias (i.e., likelihood that outcome data is missing is related both to outcome status and exposure).”

^r See Table H-8 “Serious Flaws that Would Make Epidemiological Studies Unacceptable for Use in the Hazard Assessment” under the “outcome assessment domain” “Outcome measurement or characterization” metric (page 232) which specified STROBE guideline #15 to assess this metric.

^s 15 USC §2625 (h)

EPA's "exposure characterization" domain for human studies includes the level of exposure as a fatal flaw, stating: "For all study types: The **levels** of exposure are not sufficient or adequate (as defined above)^t to detect an effect of exposure (Cooper et al., 2016)." Unlike human experimental studies, which are largely precluded for ethical reasons, human observational studies can only be based on what exposures actually occur in the real world. EPA offers no explanation of how one could know whether the levels would be "sufficient or adequate" enough to detect an effect. Given the vagaries of this metric, it could be reasonably anticipated that it would permit EPA to arbitrarily exclude quality research from its decision-making.

We recommend: EPA should not use a quantitative scoring method to assess quality in individual studies; it should not conflate study reporting with study quality; and it should not exclude otherwise quality research based on a single reporting or methodological limitation. Rather EPA should employ a scientifically valid method to assess risk of bias of individual studies.

^t EPA "as defined above" is unclear, presumably "as defined above" refers to the definition of the domain in Table H-2 page 223, "Evaluation of exposure assessment methodology that includes consideration of methodological quality, sensitivity, and validation of the methods used, degree of variation in participants, and an established time order between exposure and outcome."

3. EPA's TSCA systematic review framework does not consider financial conflicts of interest as a potential source of bias in research.

As observed by the Deputy Editor (West) of JAMA in 2010, "the biggest threat to [scientific] integrity [is] financial conflicts of interest" (39). Yet EPA's systematic review framework is silent on how it will take into account this empirically documented influence on the results of scientific research. Underscoring this EPA SR framework deficiency is the fact that recent studies empirically document that industry sponsorship produces research that is favorable to the sponsor (40, 41). The influence of financial ties on research can be traced to a variety of types of biases, and this conflict of interest needs to be distinguished from non-financial interests in the research, which can also affect research (42).

The fact that funding source needs to be accounted for in some manner is empirically supported and not a subject of scientific debate; what scientists differ on is *how* to best address funding as a potential source of bias (43, 44); for example, whether funding source is assessed as a specific risk of bias domain (43) or considered at multiple points in the evaluation (20, 44). For example, funding source is recommended as a factor to consider when evaluating risk of bias of individual studies for selective reporting, and then again for evaluating the body of evidence for publication bias, (45) and/or to be considered as a potential factor to explain apparent inconsistency within a body of evidence (14).

A 2017 Cochrane systematic review of industry sponsorship and research outcome concluded ... "industry sponsorship should be treated as bias-inducing and industry bias should be treated as a separate domain" (40). The National Academy of Sciences in its review of the EPA IRIS program's SR method found that "Funding sources should be considered in the risk-of-bias assessment conducted for systematic reviews that are part of an IRIS assessment (17)(p 79).

Notably, EPA's exclusion of consideration of funding source and other potential conflicts of interests is also internally inconsistent with EPA's own improper reliance on STROBE guidelines as quality measures: STROBE guidelines item #22 specified that "the source of funding and the role of funders, could be addressed in an appendix or in the methods section of the article" (32).

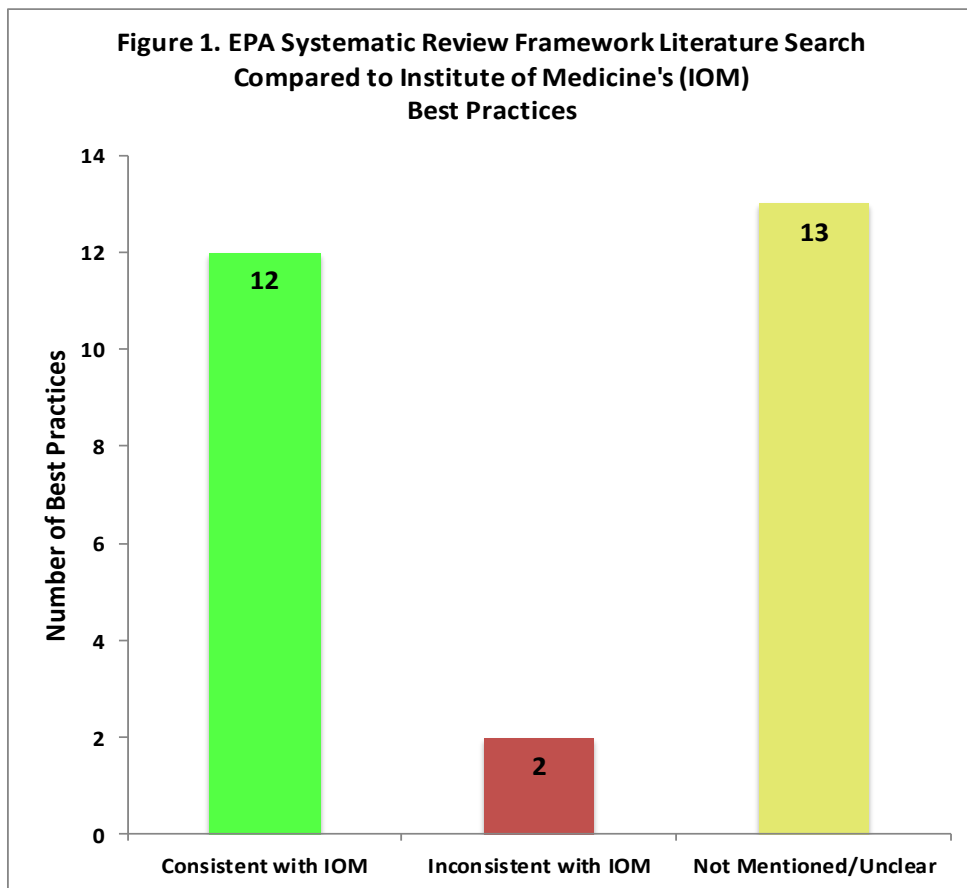
Importantly, including funding as a risk of bias as a domain does not mean excluding industry sponsored studies from EPA's hazard and risk assessment; it only means documenting funding as one of many domains of potential bias and evaluating its impact on the overall quality of the body of evidence.

We recommend: EPA should assess study and author funding source as a risk of bias domain for individual studies.

4. The literature review step of EPA’s TSCA systematic review framework incorporates select best practices, but also falls short of, or is unclear about, many other best practices for conducting a systematic and transparent literature review.

Overall, we commend the EPA for its efforts to incorporate many best practices for a comprehensive literature search in its systematic review framework. We compared EPA’s framework for systematic review to the Institute of Medicine’s (IOM’s) best practices for the literature review step of a systematic review (16)(See IOM 2011 Chapter 3. and TABLE E-1), which was applied by the National Academy of Sciences in its review of EPA’s IRIS Program methods for systematic review (17)(See Table 4-1 Pp. 43-55).

We found EPA’s framework to be consistent with 12 of IOM’s 27 best practices for conducting a literature search (Figure 1 and Appendix 1). There are two key features of EPA’s framework that are clearly inconsistent with IOM’s best practices. EPA fails: (1) to include or exclude studies based on the protocol’s pre-specified criteria, a practice that is critical to avoiding results-based decisions;^u and (2) to use two or more members of the review team, working independently, to screen and select studies, which is an essential quality-assurance measure.^v



^u See our Comment #1 regarding the EPA framework’s lack of a pre-defined protocol.

^v EPA’s framework, “Summary of the Title/Abstract Screening Conducted for the First Ten TSCA Risk Evaluations” (page 24) states that only one screener conducted the screening and categorization of titles and abstracts.

For the remaining 13 IOM best practices, EPA's framework is either unclearly stated (N=7) or the practice is not mentioned at all (N=6). However, based on the literature review methods presented in the First Ten TSCA Risk Evaluations, EPA's framework appears to have incorporated six additional best practices that are either unclear or not mentioned in EPA's SR framework: (1) work with a librarian or other information specialist trained in performing systematic reviews to plan the search strategy (IOM 3.1.1); (2) Design the search strategy to address each key research question (IOM 3.1.2); (3) Search regional bibliographic databases if other databases are unlikely to provide all relevant evidence (IOM 3.1.9); (4) Conduct a web search (IOM 3.2.5); and (5) Provide a line-by-line description of the search strategy, including the date of search for each database, web browser, etc. (IOM 3.4.1).

EPA should make its framework for conducting a literature review transparently congruent with all of IOM's best practices. This includes addressing two critical inconsistencies: (1) include or exclude studies based on the protocol's pre-specified criteria to prevent results-based decisions; and (2) Use two or more members of the review team, working independently, to screen and select studies, to ensure quality assurance. The transparency of the framework would be improved by specifying how EPA is addressing each best practice; at this juncture, how EPA intends to specifically handle many components of its literature searches could not readily be identified.

For example, the framework is unclear about whether EPA will include papers published in languages other than English. The exclusive reliance on English-language studies may lead to under-representation of the entire body of available evidence, and studies have also suggested that language bias might lead to erroneous conclusions (46). Furthermore, when considering the inclusion or update of an existing systematic review, studies have found that language-inclusive systematic reviews (including studies in languages other than English) were of the highest quality, compared with other types of reviews (47). Online translation tools are readily available to allow screeners to quickly evaluate study abstracts for relevance, and therefore we recommend EPA to incorporate non-English language studies in their screening and not simply exclude in advance these potentially relevant papers.

Additionally, EPA's framework should explicitly include rules for determining when the list of relevant studies will be considered final i.e., "stopping rules." Newer scientific studies will inevitably continue to appear in scientific journals and it will be impossible to continually attempt to include all these studies in a chemical assessment. To meet the deadlines as mandated by the Lautenberg Amendments, EPA should state clear stopping rules in the form of deadlines or criteria for when the body of included relevant studies will be finalized for the purposes of the chemicals assessment. We also strongly encourage EPA in its stated exploration of automation and machine learning tools,^w which can help speed the production of EPA's systematic reviews.

We recommend: EPA should make its framework for conducting a literature review congruent with all of the Institute of Medicine's best practices, and explicitly include rules for when the list of relevant studies will be considered final.

^w Footnote 9 page 23 states "In addition to using DistillerSR, EPA/OPPT is exploring automation and machine learning tools for data screening and prioritization activities (e.g., SWIFT-Review, SWIFT-Active Screener, Dragon, DocTER). SWIFT is an acronym for "Sciome Workbench for Interactive Computer-Facilitated Text-mining".

5. EPA's TSCA systematic review framework correctly recognizes that mechanistic data are not required for a hazard assessment, but EPA is not clear that these data, if available, can only be used to increase, and not to decrease, confidence in a body of evidence.

EPA's TSCA framework (page 172) states that EPA will use the evaluation strategies for animal and *in vitro* toxicity data to assess the quality of mechanistic and pharmacokinetic data supporting the model, and may tailor its criteria further to evaluate new approach methodologies (NAMs). We agree with EPA that mechanistic data need to be evaluated in a manner comparable to how other streams of evidence are evaluated. Data generated by alternative test methods (such as high-throughput screening methods) are not different than any other type of *in vitro* or cell-based assay data that would be considered in a systematic review. These kinds of assays provide mechanistic data. However, in this case, as described in comment # 2 above, EPA's use of its evaluation strategies for animal and *in vitro* toxicity data would entail using a quantitative scoring method that is incompatible with the best available science in fundamental ways. EPA should employ a scientifically valid method to assess risk of bias of individual studies in *all* streams of evidence, including mechanistic data.

EPA's framework (page 172) states, "the availability of a fully elucidated mode of action (MOA) or adverse outcome pathway (AOP) is not required to conduct the human health hazard assessment for a given chemical (emphasis added)." We strongly agree with EPA that mechanistic data are not needed for a hazard assessment. In addition, EPA's framework should be explicit that mechanistic data are only used to increase confidence in a hazard assessment, and never to decrease confidence.

The National Academy of Sciences explicitly considered how mechanistic data could be utilized in a systematic review for evidence integration (19). The committee came to two conclusions. First, the same protocol for evaluating relevance and study quality must be used with mechanistic data as for any other study. For example, in the report's case study on phthalates, the committee was not able to integrate results from high-throughput assays because the cell lines used were of unknown relevance to the *in vivo* mechanism of phthalate toxicity (19)(pg.78). Second, the foundation of the hazard classification in a systematic review is the animal and human data, with the mechanistic data playing a supporting role. If mechanistic data is relevant, it can be used to upgrade a hazard classification, or increase the confidence of a finding made based on evaluation of animal and human data. A hazard classification is never made based on high-throughput or other kinds of mechanistic data alone (19)(Pp. 158-9).

We recommend: EPA should be explicit that mechanistic data can only be used to upgrade a hazard classification, or increase the confidence of a finding made based on evaluation of animal and human data, and that these data will not be used to decrease confidence in a body of evidence.

6. EPA's TSCA systematic review framework is not independent of the regulatory end user of the review.

EPA's TSCA systematic review/risk assessment process is not independent of the TSCA risk management process, a conflict that is incompatible with best scientific methods. EPA's SR framework was developed and is being implemented by the Office of Chemical Safety and Pollution Prevention (OCSPP), which is also responsible for regulating the environmental exposures under TSCA review. In contrast, other EPA chemical assessment programs such as the IRIS program are intentionally placed in a non-regulatory research arm (the Office of Research and Development), to create separation from the Agency's program office responsible for regulatory decisions. This separation supports IRIS's ability to develop impartial chemical toxicity information independent of its ultimate use by EPA's program and regional office in risk assessment and risk management decisions. The National Academy of Sciences supported this in its 2018 report, stating that: "Current best practices [for systematic reviews in other medical disciplines] recommended by the Institute of Medicine (IOM 2011) suggest that the IRIS teams involved in the systematic-review process **should be independent of those involved in regulatory decision-making** who use the products of the systematic-review teams **(emphasis added)**" (15). This same principle should also be implemented across the Agency and specifically for TSCA assessments.

We recommend: EPA's systematic reviews should be produced independently of the regulatory end user of the review.

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